



Plastic waste release caused by COVID-19 and its fate in the global ocean

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The COVID-19 pandemic has led to an increased demand for single-use plastics that intensifies pressure on an already out-of-control global plastic waste problem. While it is suspected to be large, the magnitude and fate of this pandemic-associated mismanaged plastic waste are unknown. Here, we use our MITgcm ocean plastic model to quantify the impact of the pandemic on plastic discharge. We show that 8.4 ± 1.4 million tons of pandemic-associated plastic waste have been generated from 193 countries as of August 23, 2021, with 25.9 ± 3.8 thousand tons released into the global ocean representing $1.5 \pm 0.2\%$ of the global total riverine plastic discharge. The model projects that the spatial distribution of the discharge changes rapidly in the global ocean within 3 y, with a significant portion of plastic debris landing on the beach and seabed later and a circumpolar plastic accumulation zone will be formed in the Arctic. We find hospital waste represents the bulk of the global discharge (73%), and most of the global discharge is from Asia (72%), which calls for better management of medical waste in developing countries.

plastic | ocean | MITgcm | COVID-19

Plastics have an excellent strength to weight ratio, and they are durable and inexpensive, making them the material of choice for most disposable medical tools, equipment, and packaging (1, 2). The COVID-19 pandemic has demonstrated the indispensable role of plastic in the healthcare sector and public health safety (2). As of August 23, 2021, about 212 million people worldwide have been infected with the COVID-19 virus with the most confirmed cases in the Americas (47.6%) and Asia (31.22%) followed by Europe (17.26%) (3). The surging number of inpatients and virus testing substantially increase the amount of plastic medical waste (4). To sustain the enormous demand for personal protective equipment (PPE, including face masks, gloves, and face shields), many single-use plastic (SUP) legislations have been withdrawn or postponed (2). In addition, lockdowns, social distancing, and restrictions on public gathering increase the dependency on online shopping at an unprecedented speed, the packaging material of which often contains plastics (5, 6).

Unfortunately, the treatment of plastic waste is not keeping up with the increased demand for plastic products. Pandemic epicenters in particular struggle to process the waste (7), and not all the used PPEs and packaging materials are handled or recycled (8, 9). This mismanaged plastic waste (MMPW) is then discharged into the environment, and a portion reaches the ocean (10). The released plastics can be transported over long distances in the ocean, encounter marine wildlife, and potentially lead to injury or even death (11–14). For example, a recent report estimated that 1.56 million face masks entered the oceans in 2020 (15). Earlier studies have also raised the potential problem of COVID-19 plastic pollution and its impact on marine life (16–18). Some cases of entanglement, entrapment, and ingestion of COVID-19 waste by marine organisms, even leading to death, have been reported (19, 20). The plastic debris could also facilitate species invasion and transport of

contaminants including the COVID-19 virus (21–23). Despite the potential impacts, the total amount of pandemic-associated plastic waste and its environmental and health impacts are largely unknown. Here, we estimate the amount of excess plastic released during the pandemic that enters the global ocean and its long-term fate and potential ecological risk.

Results

MMPW Generation. As of August 23, 2021, the total excess MMPW generated during the pandemic is calculated as 4.4 to 15.1 million tons (Fig. 1). We use the average of scenarios with different assumptions as our best estimate (*Methods*), which is about 8.4 ± 1.4 million tons. A dominant fraction (87.4%) of this excess waste is from hospitals, which is estimated based on the number of COVID-19 inpatients (24) and per-patient medical waste generation for each country (25). PPE usage by individuals contributes only 7.6% of the total excess wastes. Interestingly, we find that the surge in online shopping results in an increased demand for packaging material. However, we find that packaging and test kits are minor sources of plastic waste and only account for 4.7% and 0.3%, respectively.

Table 1 shows the distribution of COVID-19 cases across different continents (Asia, Europe, North America, South America, Oceania, and Africa). About 70% of COVID-19 cases are found in North and South America and Asia (Table 1). We find that MMPW generation does not follow the case distribution, as most MMPW is produced in Asia (46%), followed by

Significance

Plastic waste causes harm to marine life and has become a major global environmental concern. The recent COVID-19 pandemic has led to an increased demand for single-use plastic, intensifying pressure on this already out-of-control problem. This work shows that more than eight million tons of pandemic-associated plastic waste have been generated globally, with more than 25,000 tons entering the global ocean. Most of the plastic is from medical waste generated by hospitals that dwarfs the contribution from personal protection equipment and online-shopping package material. This poses a long-lasting problem for the ocean environment and is mainly accumulated on beaches and coastal sediments. We call for better medical waste management in pandemic epicenters, especially in developing countries.

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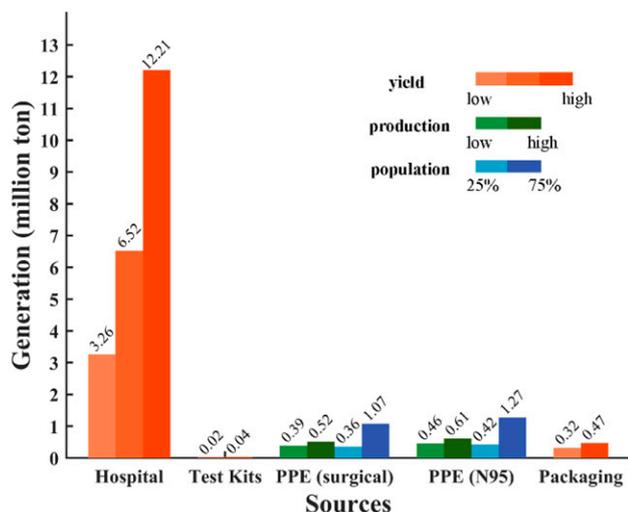


Fig. 1. Global generation of mismanaged plastics from different sources (hospital medical waste, test kits, PPE, and online packaging) attributable to the COVID-19 pandemic. High- and low-yield scenarios are considered for each source (Methods).

Europe (24%), and finally in North and South America (22%) (Table 1 and Fig. 2E). This reflects the lower treatment level of medical waste in many developing countries such as India, Brazil, and China (range between 11.5 and 76% as the low- and high-end estimates) compared with developed countries with large numbers of cases in North America and Europe (e.g., the United States and Spain) (0 to 5%) (Fig. 2A). The MMPW generated from individual PPE is even more skewed toward Asia (Fig. 2C and SI Appendix, Table S1) because of the large mask-wearing population (26). Similarly, the MMPW generated from online-shopping packaging is the highest in Asia (Fig. 2D). For instance, the top three countries in the express-delivery industry of global share are China (58%), United States (14.9%), and Japan (10.3%) followed by the United Kingdom (4%) and Germany (4%) (27).

Riverine Discharge of MMPW. Based on the MMPW production from each country and a hydrological model (28), we calculate a total discharge of 25.9 ± 3.8 (12.3 as microplastics [< 5 mm] and 13.6 as macroplastics [> 5 mm]) thousand tons of pandemic-associated plastics to the global ocean from 369 major rivers and their watersheds (Fig. 2E). We believe that the 369 rivers (account for 91% of the global riverine plastic discharge to the sea) considered here include a vast majority of the global pandemic-associated plastic discharge. The top three rivers for pandemic-associated plastic waste discharge are Shatt al Arab (5.2 thousand tons, in Asia), Indus (4.0 thousand tons, in Asia),

and Yangtze River (3.7 thousand tons, in Asia) followed by Ganges Brahmaputra (2.4 thousand tons, in Asia), Danube (1.7 thousand tons, in Europe), and Amur (1.2 thousand tons, in Asia). These findings highlight the hotspot rivers and watersheds that require special attention in plastic waste management.

Overall, the top 10 rivers account for 79% of pandemic plastic discharge, top 20 for 91%, and top 100 for 99%. About 73% of the discharge is from Asian rivers followed by Europe (11%), with minor contributions from other continents (Table 1). This pattern is different from that of the generation of MMPW (Table 1) because of the different ability of rivers to export plastic load to the ocean, which is measured as the yield ratio (defined as the ratio between the plastic discharges at the river mouth and the total MMPW generation in the watershed). The yield ratio is influenced by factors such as the distribution of plastic release along rivers and the physical conditions of rivers (e.g., water runoff and velocity) (28). The top five rivers with the highest yield ratios are the Yangtze River (0.9%), Indus (0.5%), Yellow River (0.5%), Nile (0.4%), and Ganges Brahmaputra (0.4%). These rivers have either high population density near the river mouth, large runoff, fast water velocity, or a combination of them. The combination of high pandemic-associated MMPW generations and yield ratio for Asian rivers results in their high discharge of MMPW to the ocean.

The Fate of MMPW in the Ocean. We simulate the transport and fate of the $25,900 \pm 3,800$ tons of pandemic-associated plastic waste by the Nanjing University MITgcm-Plastic model (NJU-MP) to evaluate its impact on the marine environment. The model considers the primary processes that plastics undergo in seawater: beaching, drifting, settling, biofouling/defouling, abrasion, and fragmentation (29). The model reveals that a large fraction of the river discharged plastics are transferred from the surface ocean to the beach and seabed within 3 y (Fig. 3). At the end of 2021, the mass fraction of plastics in seawater, seabed, and beach are modeled as 13%, 16%, and 71% respectively. About 3.8% of the plastics are in the surface ocean with a global mean concentration of 9.1 kg/km^2 . Our model also suggests that the discharged pandemic-associated plastics are mainly distributed in ocean regions relatively close to their sources, for example, middle- and low-latitude rivers distributed in East and South Asia, South Africa, and the Caribbean (Fig. 4 and SI Appendix, Fig. S2). The beaching and sedimentation fluxes are mainly distributed near major river mouths (Fig. 4 and SI Appendix, Fig. S2). This suggests that the short-term impact of pandemic-associated plastics is rather confined in the coastal environment.

The model suggests the impact could expand to the open ocean in 3 to 4 y. The mass fraction of plastics in the seawater is predicted to decrease in the future while those in seabed and beach are modeled to gradually increase. At the end of 2022, the fractions of riverine discharged, pandemic-associated MMPW in seawater, seabed, and beach are modeled as 5%, 19%, and 76%, respectively, and the mean surface ocean

Table 1. Percentage of the confirmed COVID-19 cases (as of August 23, 2021), the generated mass of pandemic-associated MMPW ending up in the environment, and the pandemic-associated MMPW that is transported to river mouths for different continents

Continents	Confirmed patient cases (%)	MMPW (%)	The riverine discharge (%)		
			Total	Microplastic	Macroplastic
Africa	3.8	7.9	5.9	5.6	6.1
Asia	31.2	46.3	72.5	75.6	69.6
Europe	25.7	23.8	11.2	10.0	12.5
North America	21.9	5.6	1.9	1.5	2.3
South America	17.3	16.4	6.9	6.0	7.6
Oceania	0.1	<0.1	1.6	1.3	1.9

All the percentages are of the corresponding global total.

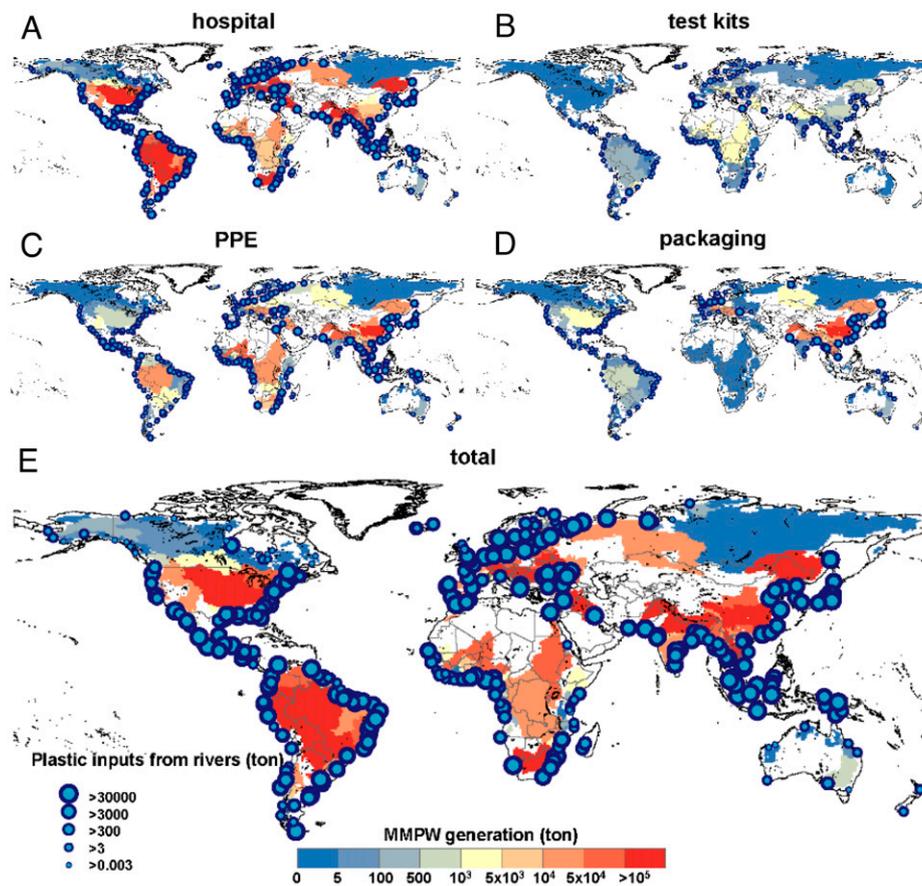


Fig. 2. Accumulated riverine discharge of pandemic-associated mismanaged plastics to the global ocean. Panels are for the discharges caused by (A) hospital medical waste, (B) COVID-19 virus test kits, (C) PPE, (D) online-shopping packaging material, and (E) the total of them. The background color represents the generated MMPW in each watershed, while the sizes of the blue circles are for the discharges at river mouths.

concentration sharply decreases to $3.1\text{kg}/\text{km}^2$. In 2025, five garbage patches in the center of subtropic gyres merge, including the four in North and South Atlantic and Pacific and the one in the Indian Ocean (Fig. 4 and *SI Appendix, Fig. S2*). Hot spots

for sedimentation fluxes are also modeled in the high-latitude North Atlantic and the Arctic Ocean in 2025 (Fig. 4 and *SI Appendix, Fig. S2*), reflecting the large-scale vertical movement of the seawaters (*SI Appendix, Fig. S3*).

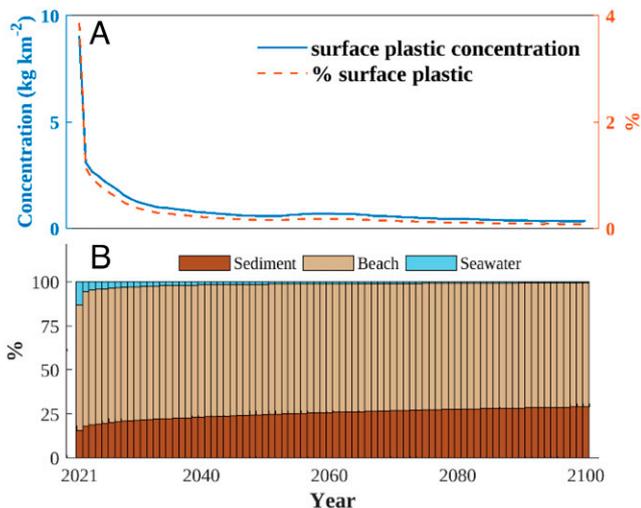


Fig. 3. Projection of the fate of discharged pandemic-associated plastics (including both microplastics and macroplastics) in the global ocean. (A) The mass fractions and average concentrations in the surface ocean. (B) The mass fractions in the seawater, seabed, and beaches.

We find a long-lasting impact of the pandemic-associated waste release in the global ocean. At the end of this century, the model suggests that almost all the pandemic-associated plastics end up in either the seabed (28.8%) or beaches (70.5%), potentially hurting the benthic ecosystems. The global mean pandemic-associated plastic concentrations in the surface ocean are predicted to decrease to $0.3\text{kg}/\text{km}^2$ in 2100, accounting for 0.03% of the total discharged plastic mass. However, two garbage patches are still modeled over the northeast Pacific and the southeast Indian Ocean, exerting persistent risk for ecosystems over there. The fate of microplastics and macroplastics are similar but with a higher fraction of macroplastics ending up in the beaches due to their lower mobility (Fig. 4 and *SI Appendix, Fig. S1*).

The Arctic Ocean appears to be a dead-end for plastic debris transport due to the northern branch of the thermohaline circulation (30). About 80% of the plastic debris discharged into the Arctic Ocean will sink quickly, and a circumpolar plastic accumulation zone is modeled to form by 2025. In this year, the Arctic seabed accounts for 13% of the global plastic sedimentation flux, but this fraction will increase to 17% in 2100. The Arctic ecosystem is considered to be particularly vulnerable due to the harsh environment and high sensitivity to climate change (31, 32), which makes the potential ecological impact of exposure to the projected accumulated Arctic plastics of special concern.

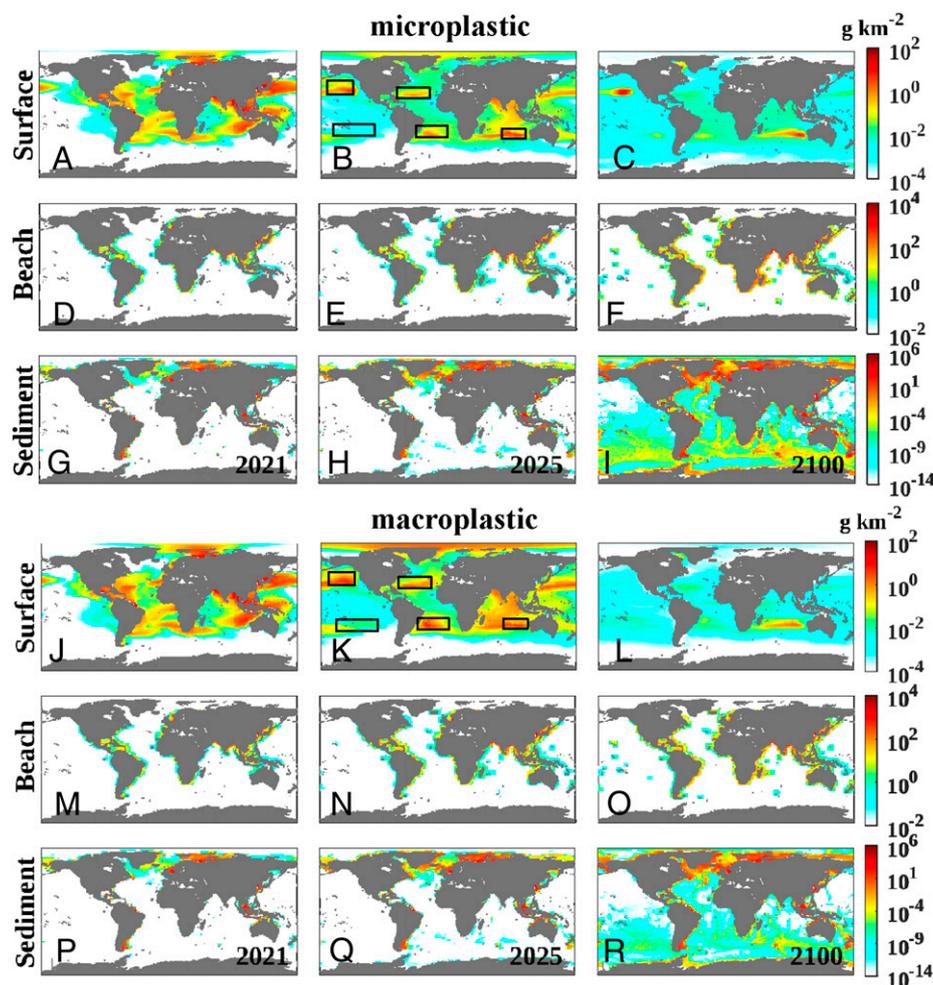


Fig. 4. Modeled spatial distribution of mass concentrations of COVID-19-associated plastics in the surface ocean (A–C, J–L), on the beaches (D–F, M–O), and the seabed (G–I, P–R) in 2021, 2025, and 2100, respectively. The black boxes on the *Top* panel indicate the five subtropical ocean gyres (North Pacific Gyre, North Atlantic Gyre, South Pacific Gyre, South Atlantic Gyre, and Indian Gyre). Panels A–I are for the microplastics, while J–R are for the macroplastics.

Discussion

It is speculated that the pandemic will not be completely controlled in a couple of years, and many of the containing policies will continue to be implemented (33). By the end of 2021, it is conservatively estimated that the number of confirmed cases will reach 280 million (34). The generated pandemic-associated MMPW will reach a total of 11 million tons, resulting in a global riverine discharge of 34,000 tons to the ocean. The MMPW generation and discharge are expected to be more skewed toward Asia due to record-breaking confirmed cases in India (3). Given the linearity between the discharge and ocean plastic mass, the fate and transport of the newly generated plastic discharge can be deduced from our current results.

There are substantial uncertainties associated with our estimate of pandemic-associated MMPW release due to the lack of accurate data (e.g., the number of used masks and online-shopping packages and the fraction of mismanaged waste under the over-capacity conditions). For example, our estimate for the discharge from face mask usage is much lower than that of Chowdhury et al. (35), which assumes that a person uses a single mask daily while we assume a mask lasts for 6 d based on survey data (*Methods*). We thus consider multiple scenarios to cap the actual situations (*Methods*). The estimated MMPW as hospital medical waste varies by $\pm 53\%$, while that from packaging and PPE vary by $\pm 25\%$ and a factor of ~ 3.5 , respectively.

The estimated amounts of riverine MMPW discharge to the ocean have also uncertainty as they are based on a coarse resolution (i.e., watershed-wise) hydrological model (28). In addition, factors such as the fragmentation, abrasion, and beaching rate of plastics in NJU-MP also have a substantial influence on the simulation results (29). Despite these uncertainties, the spatial pattern of the pandemic-associated releases and their relative fate in different compartments of the ocean is more robust.

The pandemic-associated plastic discharge to the ocean accounts for $1.5 \pm 0.2\%$ of the total riverine plastic discharges (28, 36). A large portion of the discharge is medical waste that also elevates the potential ecological and health risk (37) or even the spreading of the COVID-19 virus (38). This offers lessons that waste management requires structural changes. The revoking or delaying of the bans on SUPs may complicate plastic waste control after the pandemic. Globally public awareness of the environmental impact of PPE and other plastic products needs to be increased. Innovative technologies need to be promoted for better plastic waste collection, classification, treatment, and recycling, as well as the development of more environmentally friendly materials (15, 39). Better management of medical waste in epicenters, especially in developing countries, is necessary.

Methods

Total Plastic Waste Generation. We develop an inventory for the excess plastic waste generated due to the COVID-19 pandemic. We consider four categories

of sources: hospital-generated medical waste, virus testing kits, PPE used by residents, and online-shopping packages.

For hospital-generated medical waste, we estimate the amount by the number of hospitalization patients (n_H) and per-patient healthcare waste generation rates (HCWGR). The n_H is estimated based on the number of COVID-19 infections (n_i) and the global average hospitalization rate (HR) of this disease:

$$n_H = n_i \times \text{HR}. \quad [1]$$

The n_i and HR data are based on the statistics of the World Health Organization (3). The HCWGR of COVID-19 patients is approximately two times higher than that of general patients (40), which is calculated as a function of life expectancy (LE) and CO₂ emissions (CDE) based on Minoglou et al. (25):

$$\text{HCWGR} = 2 \times (0.014\text{LE} + 0.31\text{CDE}). \quad [2]$$

This relationship was developed based on the statistical data from 42 countries worldwide and can explain 85% of the variability of the HCWGR data (25). The LE data are from Roser et al. (41), and the CDE data are from Worldometer (42).

The virus testing kits-generated medical waste is estimated based on the number of conducted tests and the amount of waste generated per test. The former data are from Ritchie et al. (43) while the latter is from Cheon (44) and ShineGene (45). Depending on the specifications of the testing kits, the waste generated per test ranges 21 to 28 g/test.

For the PPE used by residents, we consider only face masks, as other items such as gloves and face shields are less commonly used. We use two ways to estimate the number of used masks: consumption-based and production-based. For the former way, we first assume an ideal condition that each person uses a new mask every 6 d (46), and we assume that the actual mask usage lies 25 to 75% of this situation. The population data are from United Nations (26). For the latter way, we assume that all masks produced are used up. The global production (P_W) is estimated based on the mask production in China (P_C), which is the largest mask producing country (54 to 72%) in the world (47):

$$P_W = P_C \div p, \quad [3]$$

where p is the share of Chinese-produced masks (47). We also consider two scenarios for the mass of waste generated by each mask (for surgical masks or N95 masks).

The online-shopping packaging (n_p) in this study refers to the excess part that is caused by lifestyle changes during the pandemic compared to the normal situation (no COVID-19 pandemic) ($n_{\text{no-covid}}$):

$$n_p = n_{\text{actual}} - n_{\text{no-covid}}, \quad [4]$$

where n_{actual} is the actual online package usages from 2020 to the first quarter of 2021 and is estimated based on the financial report of the top six e-commerce companies worldwide (Taobao, Tmall, Amazon, Jingdong, eBay, and Walmart) (48–52). The $n_{\text{no-covid}}$ is calculated based on the package numbers in 2019 and an average annual growth rate in recent years (53). The mass of generated plastic waste (m) is then estimated based on the average mass of plastics in the packaging material (m_p) (54):

$$m = n_p \times m_p. \quad [5]$$

MMPW Generation. The amount of MMPW for each source (i) can be calculated based on the waste generation rate of the above four sources ($R_{w,i}$), the fraction of plastic waste in the total waste (P_p), and the fraction of mismanagement waste in the total waste (P_m):

$$\text{MMPW} = \sum_{i=1}^4 R_{w,i}^p \times P_p^i \times P_m^i. \quad [6]$$

We consider the former two source categories as medical waste while the latter two as municipal waste. The P_m for each country is specified according to the waste type. The P_m of municipal waste is based on Schmidt et al. (28). There is no solid data for the P_m of medical waste, and we use the data of Caniato et al. (55) as a function of the economic status (56) and the level of treatment and disposal of waste for individual countries. The dataset includes two scenarios, and we consider an additional scenario that is 50% lower than the lower one to account for the uncertainty of this fraction.

MMPW Riverine Discharge. We estimate the river discharge of pandemic-associated MMPW to the ocean based on the watershed model developed by Schmidt et al. (28), which calculates the yield ratio of plastic discharge at the river mouth to the total MMPW generated in the entire corresponding watershed. We assume this ratio is the same for pandemic-associated plastic waste and other wastes. We consider a total of 369 major rivers and their watersheds in this study. We split the country-specific, pandemic-associated MMPW data to each watershed based on the amount of regular MMPW (28).

NJU-MP. The NJU-MP model has a resolution of 2° latitude × 2.5° longitude horizontally with 22 vertical levels and is driven by ocean physics from the Integrated Global Systems Model with 4-h time step (29). The model considers five categories of plastics with different chemical composition, and the density of each category is predetermined: polyethylene (PE, 950 kg/m³), polypropylene (PP, 900 kg/m³), polyvinyl chloride (PVC, 1,410 kg/m³), polyurethane (PU, 550 kg/m³), and others (1,050 kg m^{−3}). The plastics densities are modeled to increase when biofouled but decrease when defouled (57). The densities determine their buoyancy as low-density polymers float, whereas high-density polymers sink to the sediment (58, 59). Each category has six size bins: four belong to microplastics: <0.0781 mm, 0.0781 to 0.3125 mm, 0.3125 to 1.25 mm, and 1.25 to 5 mm, and two belong to macroplastics: 5 to 50 mm and >50 mm. There is thus a total of 60 plastic tracers in the model. We assume all the plastic debris as spheres for simplicity. The pandemic-associated MMPW discharge from rivers are released as half 5 to 10 mm and half >50 mm for macroplastics, while the largest size bin (i.e., 1.25 to 5 mm) for microplastics. After their discharge into the ocean, the plastics undergo removal by beach interception (57) and sinking to the deeper ocean and eventually on the seafloor. Biofouling of light plastic types (PE and PP) is modeled following Kooi et al. (60) but adjusted for more realistic scenarios. Three types of plastics with different degrees of biological attachment are considered. In addition, the model considers the removal processes including ultraviolet degradation, fragmentation, and abrasion.

Data Availability. The MMPW generation and river discharge data for all the countries are provided in the Environmental Biogeochemistry Modeling Group (EBMG), <https://www.ebmg.online/plastics> (61). All study data are included in the article and/or *SI Appendix*.

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1. W. Chen, L. Ciacci, N.-N. Sun, T. Yoshioka, Sustainable cycles and management of plastics: A brief review of RCR publications in 2019 and early 2020. *Resour. Conserv. Recycling* **159**, 104822 (2020).
2. N. Parashar, S. Hait, Plastics in the time of COVID-19 pandemic: Protector or polluter? *Sci. Total Environ.* **759**, 144274 (2021).
3. WHO, *WHO Coronavirus (COVID-19) Dashboard* (World Health Organization, 2021).
4. S. A. Sarkodie, P. A. Owusu, Impact of COVID-19 pandemic on waste management. *Environ. Dev. Sustain.* **23**, 7951–7960 (2020).
5. V. Thakur, Framework for PESTEL dimensions of sustainable healthcare waste management: Learnings from COVID-19 outbreak. *J. Clean. Prod.* **287**, 125562 (2020).
6. ADB, *Managing Infectious Waste During the COVID-19 Pandemic* (Asian Development Bank, 2020).
7. Institute for Global Environmental Strategies, *Waste Management During the COVID-19 Pandemic From Response to Recovery* (United Nations Environment Programme, 2020).
8. R. Picheta, *Coronavirus is Causing a Flurry of Plastic Waste. Campaigners Fear It May be Permanent* (Cable News Network, 2020).
9. UNCTAD, *Growing Plastic Pollution in Wake of COVID-19: How Trade Policy Can Help* (World Trade Organization, 2020).
10. L. C. Woodall et al., The deep sea is a major sink for microplastic debris. *R. Soc. Open Sci.* **1**, 140317 (2014).
11. C. I. Fraser et al., Antarctica's ecological isolation will be broken by storm-driven dispersal and warming. *Nat. Clim. Chang.* **8**, 704–708 (2018).
12. D. W. Laist, "Impacts of Marine Debris: Entanglement of Marine Life in Marine Debris Including a Comprehensive List of Species with Entanglement and Ingestion Records" in *Marine Debris*, J. M. Coe, D. B. Rogers, Eds. (Springer Series on Environmental Management, 1997), pp. 99–139 (1997).
13. C. L. Waller et al., Microplastics in the Antarctic marine system: An emerging area of research. *Sci. Total Environ.* **598**, 220–227 (2017).
14. J. A. Ivar do Sul, M. F. Costa, M. Barletta, F. J. A. Cysneiros, Pelagic microplastics around an archipelago of the Equatorial Atlantic. *Mar. Pollut. Bull.* **75**, 305–309 (2013).
15. T. P. Bondaroff, S. Cooke, *Masks on the Beach: The Impact of COVID-19 on Marine Plastic Pollution* (OceansAsia, 2020).
16. G. E. De-la-Torre, T. A. Aragaw, What we need to know about PPE associated with the COVID-19 pandemic in the marine environment. *Mar. Pollut. Bull.* **163**, 111879 (2021).

17. J. Canning-Clode, P. Sepúlveda, S. Almeida, J. Monterio, Will COVID-19 containment and treatment measures drive shifts in marine litter pollution? *Front. Mar. Sci.* **7**, 691 (2020).
18. M. Thiel *et al.*, COVID lessons from the global south—Face masks invading tourist beaches and recommendations for the outdoor seasons. *Sci. Total Environ.* **786**, 147486 (2021).
19. A.-F. Hiemstra, L. Rambonnet, B. Gravendeel, M. Schilthuizen, The effects of COVID-19 litter on animal life. *Anim. Biol. Leiden Neth.* **71**, 215–231 (2021).
20. H. Gallo Neto *et al.*, Mortality of a juvenile Magellanic penguin (*Spheniscus magellanicus*, Spheniscidae) associated with the ingestion of a PFF-2 protective mask during the Covid-19 pandemic. *Mar. Pollut. Bull.* **166**, 112232 (2021).
21. M. P. G. Mol, S. Caldas, Can the human coronavirus epidemic also spread through solid waste? *Waste Manag. Res.* **38**, 485–486 (2020).
22. S. C. Gall, R. C. Thompson, The impact of debris on marine life. *Mar. Pollut. Bull.* **92**, 170–179 (2015).
23. J. T. Carlton *et al.*, Tsunami-driven rafting: Transoceanic species dispersal and implications for marine biogeography. *Science* **357**, 1402–1406 (2017).
24. WHO, *Coronavirus Disease (COVID-19) Q&A: What Happens to People Who Get COVID-19?* (World Health Organization, 2020).
25. M. Minoglou, S. Gerassimidou, D. Komilis, Healthcare waste generation worldwide and its dependence on socio-economic and environmental factors. *Sustainability* **9**, 220 (2017).
26. United Nations, Department of Economic and Social Affairs, Population Division. *World Prospects 2019, Online Edition. Rev. 1* (2019). <https://population.un.org/wpp>. Accessed 29 July 2020.
27. SPBPRC, *Global Express Delivery Development Report* (State Post Bureau of The People's Republic of China, 2019). <http://www.spb.gov.cn>. Accessed 8 September 2020.
28. C. Schmidt, T. Krauth, S. Wagner, Export of plastic debris by rivers into the sea. *Environ. Sci. Technol.* **51**, 12246–12253 (2017).
29. P. Wu *et al.*, Transport and fate of all-time released plastics in the global ocean. *EarthArXiv* [Preprint] (2021). <https://eartharxiv.org/repository/view/2614/> (Accessed 14 August 2021).
30. A. Cózar *et al.*, The Arctic Ocean as a dead end for floating plastics in the North Atlantic branch of the Thermohaline Circulation. *Sci. Adv.* **3**, e1600582 (2017).
31. National Wildlife Federation, *The Arctic Environment* (National Wildlife Federation, 2010).
32. I. P. Martini *et al.*, "Northern Polar Coastal Wetlands: Development, Structure, and Land Use" in *Coastal Wetlands*, G.M.E. Perillo, E. Wolanski, D.R. Cahoon, C.S. Hopkinson, Eds. (Elsevier, 2019), pp. 153–186 (2019).
33. A. Al-Youm, *Coronavirus Pandemic May Continue Through 2022: Harvard University Researchers* (Egypt Independent, 2020).
34. IHME, *COVID-19 Projections* (Institute for Health Metrics and Evaluation, 2021).
35. H. Chowdhury, T. Chowdhury, S. M. Sait, Estimating marine plastic pollution from COVID-19 face masks in coastal regions. *Mar. Pollut. Bull.* **168**, 112419 (2021).
36. L. C.-M. Lebreton, S. D. Greer, J. C. Borrero, Numerical modelling of floating debris in the world's oceans. *Mar. Pollut. Bull.* **64**, 653–661 (2012).
37. ICRC, *Medical Waste Management* (International Committee of the Red Cross, 2011). <https://www.icrc.org/en/doc/assets/files/publications/icrc-002-4032.pdf>. Accessed 16 March 2021.
38. G. Nabi, S. Khan, Risk of COVID-19 pneumonia in aquatic mammals. *Environ. Res.* **188**, 109732 (2020).
39. N. Bellou *et al.*, Global assessment of innovative solutions to tackle marine litter. *Nat. Sustain.* **4**, 516–524 (2021).
40. G. Wei, *Medical Waste Treatment Lessons from the COVID-19 Outbreak in Wuhan* (Waste Management World, 2021). <https://waste-management-world.com/medical-waste-treatment-lessons-from-the-covid-19-outbreak-in-wuhan>. Accessed 2 September 2021.
41. M. Roser, E. Ortiz-Ospina, H. Ritchie, *Life Expectancy*. (Our World in data, 2019). <https://ourworldindata.org/life-expectancy>. Accessed 3 August 2020.
42. Worldometer, *Carbon Dioxide (CO₂) Emissions per Capita* (Worldometer, 2016). <https://www.worldometers.info/co2-emissions/co2-emissions-per-capita>. Accessed 3 August 2020.
43. H. Ritchie *et al.*, *Coronavirus (COVID-19) Testing* (Our World in data, 2020). <https://ourworldindata.org/coronavirus-testing>. Accessed 23 August 2021.
44. A. Cheon, *SoGenT DiaPlexQ Novel Coronavirus (2019-nCoV) Detection Kit (CE-IVD)* (Caremille, 2020).
45. ShineGene, *Novel Coronavirus 2019-nCoV and Influenza A/B Virus Nucleic Acid Combination Test Kit Instructions* (ShineGene Molecular Biotech Inc., 2020).
46. Dydata, *Per capita masks usage during the Chinese New Year in 2020* (Dydata, 2020). <https://www.dydata.io/datastore/detail/1854035995561955328>. Accessed 16 September 2020.
47. L. L. Thomala, *Daily production volume of medical face masks in China as of April 30, 2020, by type* (Statista, 2020).
48. Alibaba, *The Fiscal Year 2020 Annual Report and Fiscal Year 2021 Interim Report* (Alibaba Group, 2021).
49. Amazon, *2020 Annual Report and 2021 Q1 Earnings* (Amazon.com, 2021). <https://ir.aboutamazon.com>. Accessed 24 August 2021.
50. Jingdong, *Quarterly Results 2020 and Quarterly Results 2021* (JD.com, 2021). <https://ir.jd.com/quarterly-results>. Accessed 24 August 2021.
51. eBay, *Annual Reports: 2020 Annual Report and Financial Reports: 2021 Q1 Earnings Release* (eBay Inc., 2021). <https://investors.ebayinc.com/financial-information/financial-summary/default.aspx>. Accessed 24 August 2021.
52. Walmart, *Annual Report & Proxies: 2021 Annual Report and Quarterly Results: 2022 1st quarter* (Walmart, 2021). <https://stock.walmart.com/investors/financial-information/quarterly-results/default.aspx>. Accessed 24 August 2021.
53. Pitney Bowes, *Pitney Bowes Parcel Shipping Index Reports Continued Growth Bolstered by China and Emerging Markets* (BusinessWire, 2019).
54. H. Duan, G. Song, S. Qu, X. Dong, M. Xu, Post-consumer packaging waste from express delivery in China. *Resour. Conserv. Recycling* **144**, 137–143 (2019).
55. M. Caniato, T. Tudor, M. Vaccari, International governance structures for health-care waste management: A systematic review of scientific literature. *J. Environ. Manage.* **153**, 93–107 (2015).
56. The World Bank, *How does the World Bank Classify Countries?* (The World Bank) 2019).
57. D. Cressey, Bottles, bags, ropes and toothbrushes: The struggle to track ocean plastics. *Nature* **536**, 263–265 (2016).
58. M. Cole, P. Lindeque, C. Halsband, T. S. Galloway, Microplastics as contaminants in the marine environment: A review. *Mar. Pollut. Bull.* **62**, 2588–2597 (2011).
59. N. Ajith, S. Arumugam, S. Parthasarathy, S. Manupoori, S. Janakiraman, Global distribution of microplastics and its impact on marine environment—A review. *Environ. Sci. Pollut. Res. Int.* **27**, 25970–25986 (2020).
60. M. Kooi, E. H. V. Nes, M. Scheffer, A. A. Koelmans, Ups and downs in the ocean: Effects of biofouling on vertical transport of microplastics. *Environ. Sci. Technol.* **51**, 7963–7971 (2017).
61. Y. Peng, P. Wu, A. T. Schartup, Y. Zhang, The inventory of the pandemic-associated MMPW and discharge. *EBMG*. <https://www.ebmg.online/plastics>. Accessed 22 October 2021.